Inventory of Ammonia Emissions from UK Agriculture

2017

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Inventory Submission Report
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Summary

The combined UK Agriculture GHG and Ammonia emission model was used to compile the 1990-2017 ammonia emission inventory for UK agriculture, replacing the previously used National Ammonia Reduction Strategy Evaluation System (NARSES) model (spreadsheet version). The new model includes much greater sectoral, spatial and temporal resolution and also ensures consistency of approach in terms of nitrogen flows and transformations for both the ammonia and GHG emission estimates. Year-specific livestock numbers and fertiliser N use were added for 2017 and revised as appropriate for previous years. The estimate for 2017 was 244.9 kt NH_3, representing an 8.1 kt decrease from the previously reported estimate for 2016. Revisions and corrections to model parameters and historical activity data resulted in a decrease of 9.0 kt in the total estimate for 2016. Changes in activity data between 2016 and 2017 resulted in a 0.9 kt increase in emission between the two years. Ammonia emissions from agriculture have decreased by 17% over the time period 1990-2017, but have increased by 3.7% since 2005.

Table 1. Estimate of ammonia emission from UK agriculture for 2017

Source		kt NH_3 *	% of total
Livestock category			
Cattle		115.8	47
	Dairy	*57.1*	*23*
	Beef	*58.8*	*24*
Sheep†		9.7	4
Pigs		18.6	8
Poultry		37.7	15
Horses		1.2	0
Management category			
Grazing/outdoors		17.9	7
Housing		66.2	27
Hard standings		16.6	7
Manure storage		20.8	8
Manure application		61.5	25
Fertiliser application		44.9	18
Sewage sludge application		4.2	2
Digestate application		12.8	5
TOTAL		**244.9**	**100**

†Including goats and deer

* Totals may differ from sum of components due to rounding

Estimate of ammonia emission from UK agriculture for 2017

The estimate of NH_3 emission from UK agriculture for 2017 was made using the combined GHG and ammonia emission model for UK agriculture. The new model uses the same underlying approach as used in the national-scale NARSES model (Webb and Misselbrook, 2004), but incorporates a much higher level of spatial (10 km grid cells), temporal (monthly) and sectoral (greater disaggregation of dairy, beef, sheep, grassland and cropping sectors) resolution for the bottom-up calculations. As part of the model development and improvement, revisions were made to some parameters in the N-flow calculations compared with the NARSES model to ensure consistency between the estimates of ammonia and greenhouse gas emissions. Further details of the model and parametrisation are given in the UK Informative Inventory Report and National Inventory Report.

To compile the 2017 inventory of NH_3 emissions from UK agriculture, survey data were reviewed to derive livestock numbers, fertiliser use and other management practice data relevant to 2017 and to update historical activity data (1990-2016) as appropriate. Currently-used emission factors were reviewed in the light of new experimental data and amended if considered appropriate.

Key areas of revision in the 2017 inventory were:

- Correction to urea use data for 2016
- Correction to the emission factor for ammonium sulphate/diammonium phosphate type fertilisers for all years
- Correction to the energy balance equations for ‘other dairy cattle’
- Inclusion of 2017 livestock numbers
- Inclusion of 2017 N fertiliser use
- Inclusion of emissions from digestate applications to land in this report

Derivations of emission factors and reduction efficiencies assumed for mitigation practices are detailed in Appendices 1 and 2.

The estimate of emission from UK agriculture for 2017 was 244.9 kt NH_3. Cattle represent the largest livestock source and housing and land spreading the major sources in terms of manure management (Table 1). A breakdown of the estimate is given in Table 2, together with a comparison with the previously submitted 2016 inventory estimate.

Major changes between 2016 and 2017

1. Correction to urea use data for 2016

In the 1990-2015 inventory submission, the activity data for urea-N fertiliser use in the UK for 2016 was an overestimate which has been corrected in the current submission. This resulted in a substantial decrease in the emission estimate from fertilisers for 2016, but did not influence the remainder of the time series.

2. Correction to the EF for AS/DAP fertiliser types

An error whereby the EF for AS/DAP fertiliser types was not reduced if applied to non-calcareous soils was corrected in the current submission, resulting in a decrease in the estimate of emissions form fertilisers across the whole time series.

3. Correction to the energy balance equations for cattle

An error in the energy balance equations (specifically the maintenance requirement) for ‘other dairy cattle’ categories (i.e. dairy calves, replacements and heifers) was corrected. This resulted in a lower estimate of N excretion for these cattle categories across the entire time series and hence lower ammonia emission estimates.

4. 2017 livestock numbers

Headline changes from 2016 were:
Cattle – a small decrease in cattle numbers, by 0.4% for dairy cows and 0.5% for other cattle
Pigs – a 2.1% increase in pig numbers
Sheep – a 2.6% increase in sheep numbers
Poultry – a 5.3% increase in total poultry numbers, 3.5% increase in layers, 6.3% increase in broilers

5. 2017 N fertiliser use

Total fertiliser N use decreased by 1.9% from 2016 to 2017 and urea-based fertiliser N use decreased by 5.8%.

6. Inclusion of emissions from digestate application to land

Applications of digestate arising from the anaerobic digestion of food waste and purpose-grown crops was not previously included in this report, although was included in the officially reported UK total for agriculture (NAEI website: http://naei.beis.gov.uk/data/). This source has now been included in this stand-alone agriculture sector report. In addition, livestock manure going through the anaerobic digestion process are now also accounted, with the additional emission from the application of digested manure (above that from undigested manure application) being included in the estimate of ‘Digestate application’. This has been revised across the entire time series.

The anaerobic digestion of livestock manure is assumed to result in a digestate which has 80% of the total N content in the ammoniacal form. The emission factor for digestate applications to land is given by Nicholson et al. (2017) as 42% of applied total N, which equates to 52.5% of TAN (assuming 80% TN as TAN). A further assumption is that all digestate is applied using a band spreading technique, with a 30% reduction in emission.

Table 2. Estimate of ammonia emissions (kt NH_3) from UK agriculture, 2017*

Source	2016 as per 2018 submission	**2016 as per 2019 submission**	**Reasons for change between submissions**	**2017**	**Reasons for change from 2016**
Cattle			Correction to the energy balance equations for 'other dairy cattle' resulting in lower N excretion estimates		Small decrease in cattle numbers, but increase in dairy cow milk yield
Grazing	9.4	**8.7**		**8.7**	
Landspreading	35.7	**34.5**		**35.4**	
Housing	42.8	**41.5**		**41.6**	
Hard standings	17.2	**16.7**		**16.7**	
Storage	13.7	**13.3**		**13.4**	
Total Cattle	119.0	**114.7**		**115.8**	
Sheep[†]			No change		An increase in sheep numbers
Grazing	6.5	**6.5**		**6.7**	
Landspreading	1.0	**1.0**		**1.1**	
Housing	1.1	**1.1**		**1.2**	
Storage	0.7	**0.7**		**0.8**	
Total Sheep	9.3	**9.3**		**9.6**	
Horses	1.3	**1.3**	No change	**1.2**	A reduction in horse numbers
Pigs			No change		An increase in total pig numbers.
Outdoor	1.2	**1.2**		**1.2**	
Landspreading	4.1	**4.1**		**4.2**	
Housing	9.9	**9.9**		**10.1**	
Storage	3.1	**3.1**		**3.2**	
Total Pigs	18.3	**18.3**		**18.6**	
Poultry			Correction to N excretion value for laying hens		Increase in total poultry numbers.
Outdoor	0.9	**0.9**		**0.9**	
Landspreading	19.6	**19.9**		**20.5**	
Housing	12.1	**12.9**		**12.9**	
Storage	3.1	**3.3**		**3.3**	
Total Poultry	35.6	**36.5**		**37.7**	
Fertiliser	56.1	**47.8**	Correction to urea use activity data for 2016; correction to the emission factor for AS/DAP type fertilisers	**44.9**	Reduction in total fertiliser N use and in the proportion applied as urea
Sewage sludge	**4.2**	4.2		4.2	
Digestate	**9.1**	11.7	Inclusion of additional emission from livestock manure digestate	12.8	Increased quantity of digestate applied to land
TOTAL	**253.0**	**244.0**		**244.9**	

*Totals may differ from sum of components due to rounding
†Including goats and deer

Emission Trends: 1990 - 2017

Retrospective calculations based on the most recent inventory methodology were made for the years 1990 to 2017 (Table 3). There has been a steady decline in emissions from UK agriculture over the period 1990 – 2010, largely due to declining livestock numbers (Fig. 1) and fertiliser N use (Fig. 2), but also from increases in production efficiency, but this decline has levelled off in recent years. Emissions have declined by 17% since 1990, but increased by 3.7% since 2005, due in part to increases in urea fertiliser use, livestock numbers and particularly to increasing quantities of digestate applied to land from anaerobic digestion of food-waste, crops and livestock manure.

Table 3. Estimates of ammonia emission from UK agriculture 1990 – 2017

Source	1990	2000	2005	2010	2015	2017
Total	**296.3**	**252.2**	**236.1**	**222.2**	**239.0**	**244.9**
Cattle	123.5	118.5	117.4	113.9	113.5	115.8
Sheep	12.0	11.7	9.8	8.5	9.5	9.6
Pigs	40.5	30.5	21.5	17.2	18.0	18.6
Poultry	52.3	50.5	42.4	34.6	35.8	37.7
Horses	1.0	1.4	1.6	1.5	1.3	1.2
Fertiliser	58.9	37.8	39.2	41.0	47.3	44.9
Sewage sludge	1.5	1.7	3.6	3.7	4.2	4.2
Digestate	0.0	0.0	0.4	1.6	9.2	12.8

Figure 1. Trends in livestock numbers 1990 – 2017. Changes are relative to a reference value of 100 in 1990.

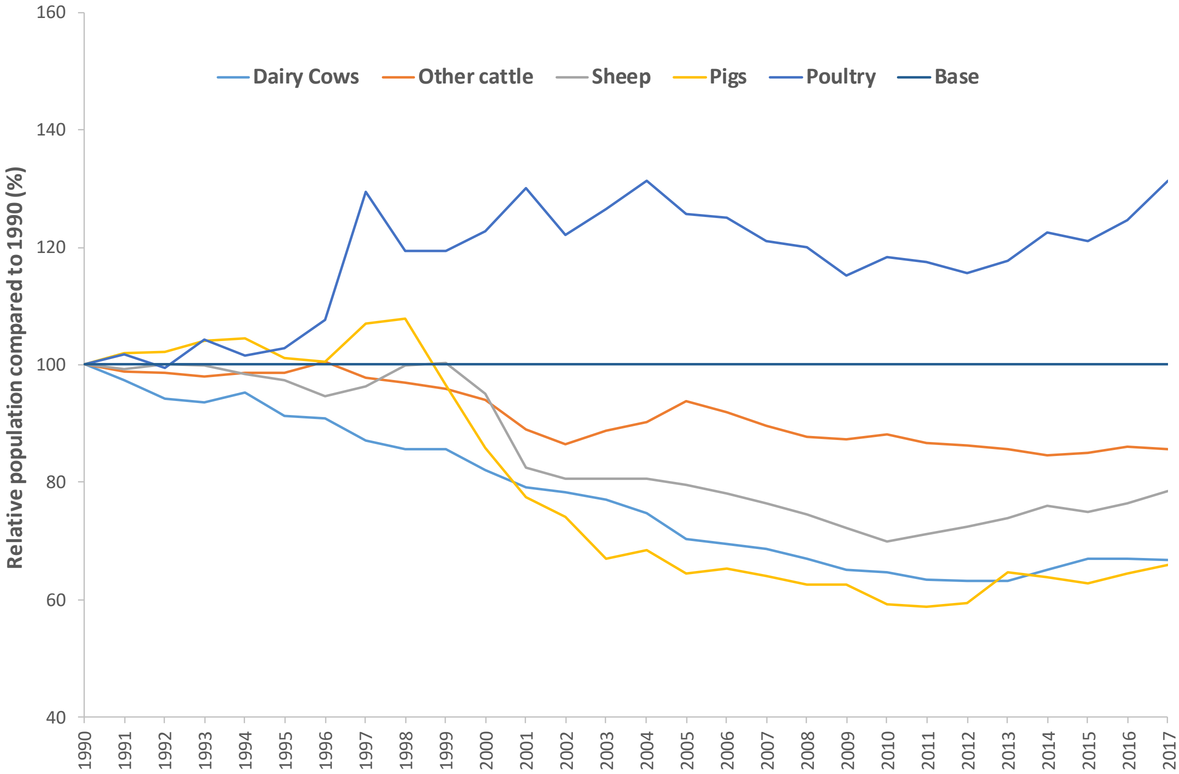
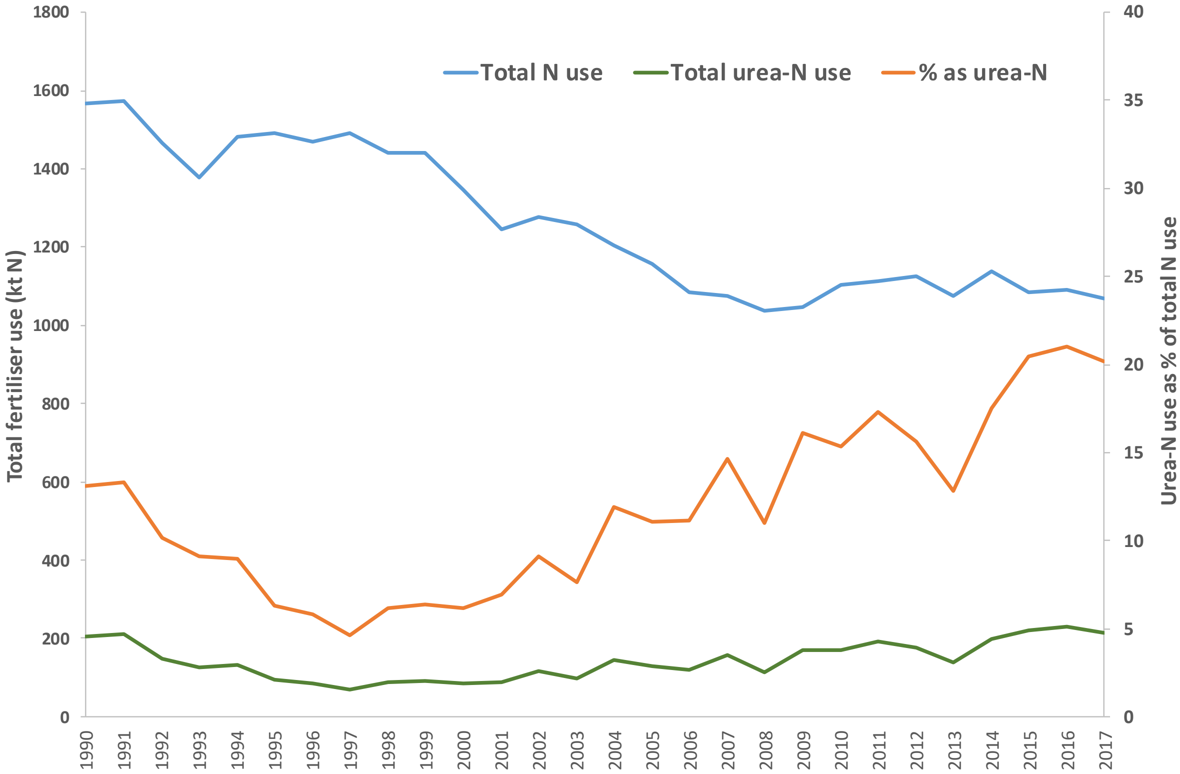


Figure 2. Changes in fertiliser N use 1990 – 2017.



Uncertainties

An estimate of the uncertainties in the emission inventory estimate was conducted using Monte Carlo simulation, in which a probability distribution function was provided for each of the model inputs (activity or emission factor data), based on the distribution of raw data or, where no or only single estimates exist, on expert assumptions. The 95% confidence interval for the total inventory estimate was estimated to be approximately ±15% (i.e. ±36.7 kt NH_3 for the 2017 estimate).

NB: uncertainties related to emissions form goats, deer, horses and sewage sludge and digestate applications to land are not currently included in this overall estimate.

Appendix 1:
Ammonia Emission Factors for UK Agriculture

Introduction

This report described the emission factors (EFs) and where appropriate standard errors (SE) for ammonia (NH_3) emissions from agricultural sources that are to be used in the improved greenhouse gas (GHG) emission inventory for UK agriculture being developed under the UK government-funded Defra project AC0114. The improved GHG inventory for UK agriculture will use a nitrogen (N) mass flow approach in calculating emissions from livestock manure management with the initial N input as excretion by livestock and subsequent losses and transformations (between organic and total ammoniacal N, TAN) being modelled at each management stage i.e. livestock housing, manure storage/treatment and manure application to land. Ammonia EFs are expressed as a percentage of the TAN content of the manure N pool at each management stage. In addition, EF are described for emissions from grazing returns (expressed as a percentage of TAN, which is generally equated with the urine fraction of the excreta) and for N fertiliser applications (with the EF expressed as a percentage of the total fertiliser N). Country- and practice-specific EFs have been derived for the major emission sources across the different agricultural sectors as described below.

1. Livestock housing

1.1. Cattle

Emission factors for two types of cattle housing are currently defined; slurry systems (solid-floor, cubicle housing with scraped passage) and deep litter straw-bedded housing generating farmyard manure (FYM). There is no differentiation between dairy and beef cattle, but a different EF was derived for calves on deep litter based on limited measurement data and the assumption that the straw bedding to excreta ratio is much greater for calves than for older cattle (Table 1). The underlying studies from which these EFs are derived are given in Annex 1 (Table A1).

It is recognised that slatted-floor slurry systems also exist for dairy and beef systems, particularly in Northern Ireland and Scotland, and that the current slurry housing system EF is not representative of these systems. Emission measurements being undertaken on such systems in the Republic of Ireland may provide useful data from which the UK can derive a system-specific EF.

Table 1. Cattle housing EFs (as % of TAN deposited in the house)

Housing system	EF	SE	n
Slurry, all cattle	27.7	3.85	14
Deep litter (FYM), all cattle except calves	16.8	1.97	10
Deep litter (FYM), calves	4.2	1.62	2

Seasonal differentiation in the EF is not included in the inventory. The EF for housing might be expected to be greater in summer, because of higher temperatures. However, work by Phillips *et al*. (1998) showed that summer emissions from dairy cattle housing, where the cattle come in for part of the day for milking, were of a similar magnitude to winter emissions. Further measurements have been conducted on modern dairy cow year-round housing units under Defra project AC0123 which will further inform the inventory in this area.

1.2. Pigs

As for cattle, housing EFs for pigs have been derived for two management systems, slurry-based and FYM-based, but for a larger number of animal categories (Table 2). A review conducted as part of Defra project AC0123 in 2012 concluded that pig housing has not changed considerably over the inventory reporting period and that the EF reported here are relevant for current housing systems. However, this should be kept under regular review as the Industrial Emissions Directive (previously Integrated Pollution Prevention and Control) and its requirement for large producers to comply with Best Available Techniques for minimising emissions should mean that there is a shift over time towards lower emission housing systems (this may be reflected in uptake of specific mitigation options rather than systemic differences in housing design).

Table 2. Pig housing EFs (as % of TAN deposited in the house)

Housing system	EF	SE	n
Dry sows on slats	22.9	14.9	2
Dry sows on straw	43.9	9.62	12
Farrowing sows on slats	30.8	2.96	7
Farrowing sows on straw	43.9	dry sows value used	
Boars on straw	43.9	dry sows value used	
Finishing pigs on slats	29.4	2.27	17
Finishing pigs on straw	26.6	5.11	15
Weaners on slats	7.9	2.01	2
Weaners on straw	7.2	based on weaners on slats value	

Most measurements have been made for finishing pigs on either slatted floor or straw-bedded systems, with fewer or no measurements for the other pig categories (Table A2).

1.3. Poultry

Measurements have been made from poultry housing for the poultry categories laying hens, broilers and turkeys (Table A3). For pullets, breeding hens and other classes of poultry not categorised in the table above, a weighted average of the broiler and turkey data were used to derive an emission factor of 14.1%. Laying hen systems are further categorised as cages without belt-cleaning, perchery, free-range and cages with belt cleaning. Of these, the cages without belt cleaning, perchery and the housing component of free-range systems are all classified as ‘deep pit’ with a common EF. There are currently no measurements for more recent ‘enriched cage’ systems, although Defra project AC0123 will report on these.

Table 3. Poultry housing EFs (as % of TAN deposited in the house)

Housing system	EF	SE	n
Layers, deep pit (cages, perchery, free-range)	35.6	8.14	7
Layers, cages with belt-cleaning	14.5	4.79	5
Broilers	9.9	0.93	15
Turkeys	36.2	30.53	3
Pullets, breeding hens and all other poultry	14.1	Based on broilers and turkeys	

1.4. Sheep

No specific measurements have been conducted for sheep housing, so the same value is used as for straw-bedded cattle housing i.e. 16.8% of the TAN deposited in the house.

1.5. Horses

Horses kept on agricultural holdings have an assumed N excretion of 50 kg per animal per year and are assumed to spend 25% of the year housed. Emission factors (expressed as %TAN) are assumed to be the same as for cattle on FYM.

2. Hard standings (unroofed outdoor concrete yards)

2.1. Cattle

Based on Misselbrook et al. (2006) an EF of 75% of the TAN left after scraping is assumed, based on mean measured values of 0.47 and 0.98 g NH_3-N animal^{-1} h^{-1} for dairy and beef cattle, respectively, with respective standard errors of 0.09 (n=28) and 0.39 (n=30) g NH_3-N animal^{-1} h^{-1}.

2.2. Sheep

An EF of 75% of the TAN left after scraping is also assumed for sheep, based on Misselbrook et al. (2006) and measured mean value of 0.13 g NH_3-N animal^{-1} h^{-1} and a standard error of 0.09 (n=7) g NH_3-N animal^{-1} h^{-1}.

3. Manure storage

3.1. Slurry

Derived EF for cattle and pig slurry storage are given in Table 4. Measurements from slurry lagoons and above-ground tanks are generally reported as emission per unit area, with only few studies containing sufficient information from which to derive an EF expressed as a percentage of the TAN present in the store (Tables A4 and A5). The EF for lagoons, in particular, are high and substantiated by very little underlying evidence (with no differentiation between pig and cattle slurries) so further measurements are warranted for this source. Emissions from below-slat slurry storage inside animal housing are assumed to be included in the animal housing EF, so below-slat storage does not appear as a separate storage category. As only few measurement data are available for EF derivation, and some categories of storage 'read across' from others, a default uncertainty estimate of ±30% for the 95% confidence interval is suggested for all slurry storage categories.

Table 4. Slurry storage EF (as % of TAN present in the store)

Storage system	EF	Uncertainty (95% CI)
Cattle slurry above-ground store (no crust)	10†	3.0
Cattle slurry weeping wall	5	1.5
Cattle slurry lagoon (no crust)	52	15.6
Cattle slurry below-ground tank	5‡	1.5
Pig slurry above-ground store	13	3.9
Pig slurry lagoon	52	15.6
Pig slurry below-ground tank	7*	2.1

†assumed to be double that of crusted slurry (for which measurements were made); ‡assumed to be the same as for above-ground slurry store with crust; *assumed to be half the value of above-ground slurry store

3.2. Solid manure

Derived EF for cattle, pig and sheep FYM and poultry manure storage are given in Table 5. There is large variability in the EF for cattle and pig FYM, with weather conditions in particular influencing emissions, and a combined EF of 28.2% (SE 6.28) is probably justified. Details of the underlying data are given in Tables A4, A5 and A6. The EF for horse FYM is assumed to be the same as that for cattle FYM.

Table 5. FYM and poultry manure storage EF (as % of TAN present in the store)

Storage system	EF	SE	n
Cattle FYM	26.3	8.28	10
Pig FYM	31.5	10.33	6
Sheep FYM	26.3	Cattle FYM EF used	
Layer manure	14.2	2.99	8
Broiler litter	9.6	2.69	11
Other poultry litter (excluding ducks)	9.6	Broiler litter EF used	
Duck manure	26.3	Cattle FYM EF used	

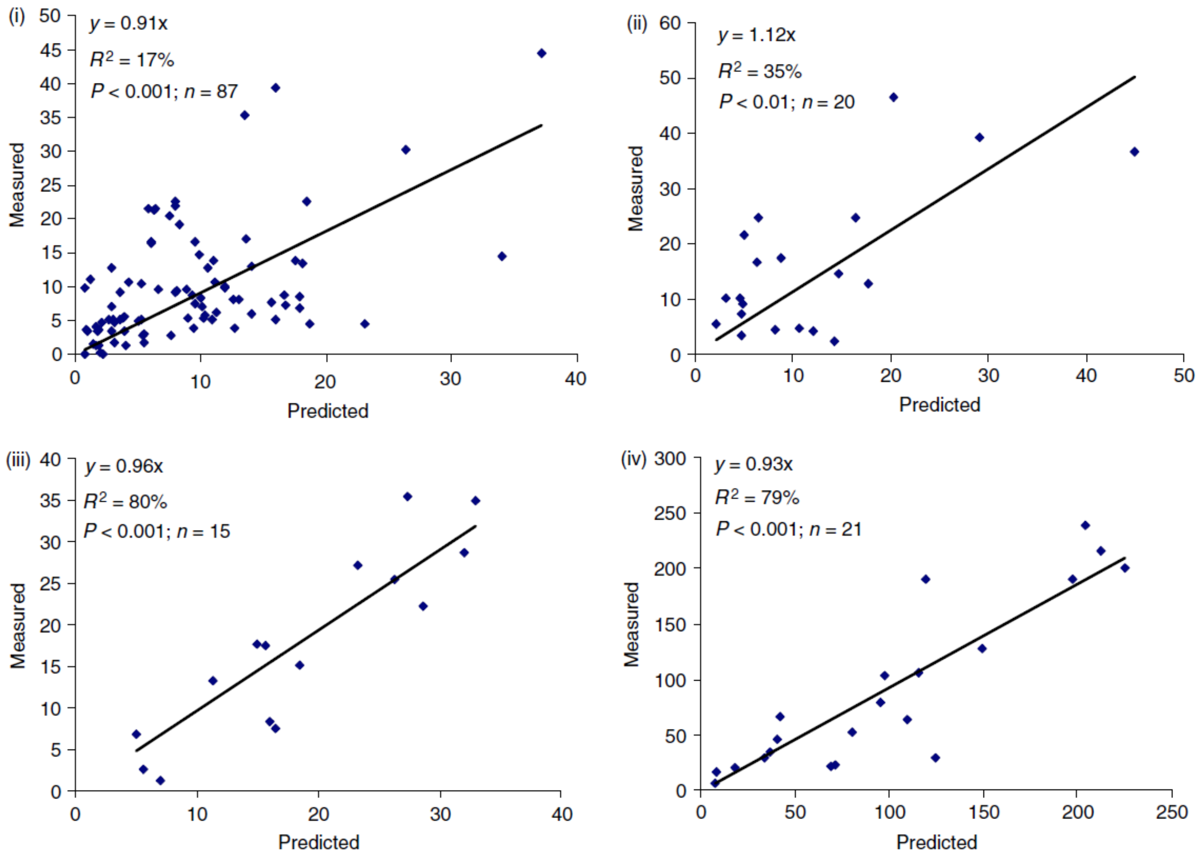
4. Manure application

Emission factors following manure applications to land are derived using the MANNER_NPK model (Nicholson et al., 2013), which established standard emission functions using a Michaelis-Menten curve fitting approach for different manure types and applied modifiers according to soil moisture, land use and slurry dry matter content (Table 6). Other modifiers included in the model according to wind speed and rainfall within 6 hours of application were not included in the national scale derivation of EF. Modifiers according to application method (splashplate assumed as baseline) and timing of soil incorporation are included as mitigation methods associated with an emission reduction efficiency and are detailed in the separate report on NH_3 emission mitigation techniques. Table 7 shows the resulting EF as used in the national inventory. Uncertainties for the weighted average EF in Table 7 were derived from the error terms in the modelled vs. observed plots using the MANNER_NPK model against UK-specific available data for cattle slurry, pig slurry, FYM (cattle and pig) and poultry manure (Fig. 1).

Table 6. Ammonia EF and modifiers according to the MANNER_NPK model

Manure type	Standard EF (as % of TAN applied)	Soil moisture modifier	Land use modifier	Slurry DM modifier	
				Slope	Intercept
Cattle slurry	32.4	x1.3 for dry soil (summer, May-July); x0.7 for moist soil	x0.85 for arable; x1.15 for grassland	8.3	50.2
Pig slurry	25.5	-	-	12.3	50.8
FYM (incl. duck)	68.3	-	-	-	-
Poultry manure	52.3	-	-	-	-

Figure 1. MANNER_NPK model performance against UK data sets for ammonia emissions following land spreading (Nicholson et al., 2013). Cattle slurry (I), pig slurry (II), FYM (III) and poultry manure (IV).



Standard errors for the derived slope values were 0.073, 0.148, 0.061 and 0.063 for I, II, III and IV, respectively.

Table 7. Manure application EF (as % of TAN applied to land)

Manure type	Land use	Season	Slurry DM	EF, %TAN	95% confidence interval, %TAN
Cattle slurry	Grassland	Summer	<4%	32.4	
			4-8%	48.4	
			>8%	64.5	
			Weighted average	52.5	8.4
Cattle slurry	Grassland	Rest of year	<4%	17.4	
			4-8%	26.1	
			>8%	34.7	
			Weighted average	28.2	4.5
Cattle slurry	Arable	Summer	<4%	23.9	
			4-8%	35.8	
			>8%	47.7	
			Weighted average	38.8	6.2
Cattle slurry	Arable	Rest of year	<4%	12.9	
			4-8%	19.3	
			>8%	25.7	
			Weighted average	20.9	3.4
Pig slurry	-	-	<4%	19.2	
			4-8%	31.8	
			>8%	44.3	
			Weighted average	24.2	6.4
FYM (all)	-	-	-	68.3	8.7
Poultry manure (all)	-	-	-	52.3	7.1

5. Grazing and outdoor livestock

5.1. Cattle and sheep

The average EF for cattle and sheep (there was no evidence to warrant differentiation) was derived from a number of grazing studies (Table A7) with a range of fertiliser N inputs to the grazed pasture. Emissions due to the fertiliser applied to the grazed pasture were discounted using a mean EF for ammonium nitrate applications to grassland (1.4% of N applied). The remaining emission was expressed as a percentage of the estimated urine N (equated here with the TAN in excreta) returned to the pasture by the grazing cattle or sheep. A mean EF of 6% of excreted TAN, with a standard error of 0.7 (n=20) was derived. This value is also assumed for grazing deer and goats.

5.2. Outdoor pigs

Only two studies have made measurements of NH_3 emissions from outdoor pigs (Table A8), and sufficient data were provided from only one of these to derive a rounded EF of 25% of TAN excreted, with an assumed 95% confidence interval of ± 7.5 % of TAN excreted.

5.3. *Outdoor poultry*

No studies of emissions from outdoor poultry have been reported. An EF of 35 % of excreted UAN has been assumed, as it is likely that emissions from freshly dropped excreta will be substantially lower than from applications of stored manure in which hydrolysis of the uric acid will have occurred to a greater extent. The 95% confidence interval for this EF is assumed to be ± 15 % of UAN excreted.

6. **Nitrogen fertiliser applications**A model based on Misselbrook et al. (2004) but modified according to data from the Defra-funded NT26 project is used to estimate EF for different fertiliser types. Each fertiliser type is associated with an EF_{max} value, which is then modified according to soil, weather and management factors (Table 8). Soil placement of N fertiliser is categorised as an abatement measure and is detailed in the separate report on NH_3 emission mitigation techniques.

Table 8. Nitrogen fertiliser application EF

Fertiliser type	EF_{max} (as % of N applied)	Modifiers†
Ammonium nitrate	1.8	None
Ammonium sulphate and diammonium phosphate	45	Soil pH
Urea	45	Application rate, rainfall, temperature
Urea ammonium nitrate	23	Application rate, rainfall, temperature
Other N compounds	1.8	None

†Modifiers:

Soil pH – if calcareous soil, assume EF as for urea; if non-calcareous, assume EF as for ammonium nitrate

Application rate

- if <=30 kg N ha^{-1}, apply a modifier of 0.62 to EF_{max}
- if >=150 kg N ha^{-1}, apply a modifier of 1 to EF_{max}
- if between 30 and 150 kg N ha^{-1}, apply a modifier of ((0.0032xrate)+0.5238)

Rainfall – a modifier is applied based on the probability of significant rainfall (>5mm within a 24h period) within 1, 2, 3, 4 or 5 days following application, with respective modifiers of 0.3, 0.5, 0.7, 0.8 and 0.9 applied to EF_{max}.

Temperature – apply a modifier, with the maximum value constrained to 1, of

$$RF_{temp} = e^{(0.1386\times(T_{month} - T_{UKannual}))} / 2$$

where $T_{UKannual}$ is the mean annual air temperature for the UK

An uncertainty bound to the EF_{max} values of ±0.3 x EF_{max} is suggested based on the measurements reported under the NT26 project.

7. **Digestate applications to land**

7.1. Food and crop-based digestates

The emission factor for land spreading of digestates from food waste sources is 1.75 kg NH_3-N t^{-1} food digestate (range 1.5 – 2 kg) (WRAP, 2016; Nicholson et al., 2017; Fiona Nicholson, ADAS, pers. comm.). The emission factor for land spreading digestates from non-manure, non-food waste materials is 0.68 kg NH_3-N t^{-1} digestate. For non-manure, non-food digestates, the latest evidence of spreading emissions (Cumby *et al.*, 2005; WRAP, 2016) was combined with an analysis of inputs to all AD sites in the UK (NNFCC, 2018) to produce an average emission factor of 1.19 kg NH_3-N t^{-1} feedstocks (range 1.07 – 1.31 kg).

7.2. Livestock manure based digestate

The emission factor for livestock manure based digestate is 42% of the applied total N (Nicholson et al., 2017). Manure digestate is assumed to have a TAN content equivalent to 80% of the total N, so the EF expressed as a proportion of the TAN (to be comparable with EF for manure applications to land) is 52.5%. A reduction factor of 30% is applied to the EF as it is assumed that all manure-based digestate is applied to land using a low emission application method (30% reduction representing band spreading).

7.3. Activity data

The amounts of materials treated in UK AD plants are considerable, and this source has been growing rapidly. Plants are listed in the database for AD sites (NNFCC, 2018) together with estimates of volume input of feedstock by type (food waste, crop, livestock manure, other). A reduction factor of 0.84 (WRAP, 2014) is applied to the input values to provide an estimate of digestate quantities, reflecting the fact that the amount of digestate produced in comparison to the amount of inputs used at the site is usually lower (due to the recycling of digestate to catalyse the process in the digester etc.). For livestock manure, types were categorised as cattle, pig, poultry, equine and miscellaneous animal. In the inventory calculations, miscellaneous animal was assumed to be cattle slurry. To estimate the quantity of N associated with the total volume of each manure type, RB209 values for typical manure N content are used: 2.6, 3.6, 24 and 7 kg t^{-1} for cattle slurry, pig slurry, poultry manure and equine manure, respectively.

Annex 1: Sources of underlying data for the UK ammonia emission factors

Table A1. Studies delivering cattle housing EF

Study	Emission g NH_3-N lu^{-1} d^{-1}	No. studies	Emission Factor % TAN	Notes on derivation of EF as %TAN
Slurry-based systems				
Demmers et al., 1997	38.6	1	31.1	Dairy cows 1995, assume N excretion of 100 kg N per year
WA0653	21.2	6	19.2	Dairy cows 1998/99, assume N excretion of 105 kg N per year
Dore et al., 2004	72.5	1	53.1	Dairy cows 1998/99, assume N excretion of 105 kg N per year
WAO632/AM110	50.8	3	39.4	Using actual N balance data
Hill, 2000	29.4	1	22.8	Dairy cows 1997, assume N excretion of 104 kg N per year
AM0102	30.5	2	23.7	Dairy cows 2003, assume N excretion of 113 kg N per year
Mean	40.5		31.6	
Weighted mean	34.3		**27.7**	
Straw-bedded systems				
WA0618 (PT)	20.6	1	18.3	Growing beef, assume N excretion of 56 kg N per year
WAO632/AM110 (PT)	35.0	3	21.6	Using actual N balance data
WA0722	33.2	1	22.9	Dairy cows, 6,500 kg milk per year, therefore assume N excretion of 112 kg N per year
AM0103 (PT)	13.9	1	11.7	Growing beef, values directly from report
AM0103 (Comm farm)	16.7	1	13.4	Dairy cows, assuming 125 g TAN excretion per day (AM0103 report)
AC0102	14.0	3	12.5	Growing beef, assume N excretion of 56 kg N per year
Mean	22.2		16.7	
Weighted mean	23.1		**16.8**	
Calves				
Demmers et al. 1997	13.0	1	5.8	Assume calf weight 140 and N excretion 38 kg N per year
Koerkamp et al. 1998	6.2	1	2.6	Assume calf weight 140 and N excretion 38 kg N per year
Mean	9.6		**4.2**	

Table A2. Studies delivering pig housing EF

Study	Emission g N $lu^{-1}d^{-1}$	No. studies	Emission Factor % TAN	Notes on derivation of EF as %TAN
Dry sows on slats				
Peirson,1995	17.0	2	22.9	Assume N excretion of 15.5kg
Dry sows on straw				
Peirson,1995	9.4	2	12.6	Assume N excretion of 15.5kg
Koerkamp et al., 1998	14.7	1	19.8	Assume N excretion of 15.5kg
OC9523	26.2	4	35.3	Assume N excretion of 15.5kg
AM0102	50.6	5	68.1	Assume N excretion of 15.5kg
Mean	25.2		34.0	
Weighted mean	15.7		**43.9**	
Farrowing sows on slats				
Peirson,1995	32.4	3	33.8	Assume N excretion of 22.5kg (1995 value)
Koerkamp et al., 1998	20.7	1	23.1	Assume N excretion 22.5kg (1995 value), live weight 240 kg
AM0102	27.0	3	30.4	Assume N excretion 15.5kg (2002/03 value)
Mean	26.7	7	29.1	
Weighted mean	20.7		**30.8**	
Farrowing sows on straw				
	Use dry sows value			
Boars on straw				
	Use dry sows value			
Finishers on slats				
Peirson, 1995	71.7	3	26.9	Assume fatteners 20-80 kg, N excretion 13.9kg (1995 value)
Demmers, 1999	105.8	1	25.3	Mean weight 25.7kg, N excretion 11.2kg (1995 value)
Koerkamp et al. 1998	51.2	1	16.7	Approx. 35 kg finishers, assume N excretion 11.2 kg (1995 value)
WA0632	79.2	4	40.4	Using actual N balance data
WA0720 (fan vent, comm farm)	103.5	1	41.5	Assume fatteners 20-80 kg, N excretion 13kg (mean of 2 weight ranges for year 2002)
WA0720 (acnv, comm farm)	77.2	3	31.0	Assume fatteners 20-80 kg, N excretion 13kg (mean of 2 weight ranges for year 2002)
WA0720 (part slat, comm farm)	51.5	2	20.7	Assume fatteners 20-80 kg, N excretion 13kg (mean of 2 weight ranges for year 2002)

Study	Emission g N $lu^{-1}d^{-1}$	No. studies	Emission Factor % TAN	Notes on derivation of EF as %TAN
WA0720 (fan vent, Terrington)	47.7	1	21.6	40-95 kg finishers, assume N excretion 15.5 kg per year
WA0720 (part slat, Terrington)	38.7	1	17.6	40-95 kg finishers, assume N excretion 15.5 kg per year
Mean	69.6	17	26.8	
Weighted mean	71.4		**29.4**	
Finishers on straw				
Peirson (1995)	54.2	2	20.3	Assume fatteners 20-80 kg, N excretion 13.9kg (1995 value)
Koerkamp et al., 1998	28.2	1	9.2	Approx. 35 kg finishers, assume N excretion 11.2 kg (1995 value)
WA0632	122.2	4	53.7	Using actual N balance data
AM0102	24.0	1	9.6	Assume fatteners 20-80 kg, N excretion 13kg (mean of 2 weight ranges for year 2002)
AM0103 Terrington	47.0	2	23.6	Values directly from report
AM0103 Commercial	34.1	1	10.9	Finishers 20-60 kg, N excretion 13kg (mean of 2 weight ranges for year 2002)
AC0102	42.0	4	16.6	Finishers 30-60 kg, N excretion 11.9kg (mean of 2 weight ranges for year 2002)
Mean	50.2	15	20.6	
Weighted mean	63.0		**26.6**	
Weaners on slats				
Peirson, 1995	34.8	1	9.9	Assume N excretion 4.4kg (1995 value)
Koerkamp et al. 1998	20.7	1	5.9	Assume N excretion 4.4kg (1995 value)
Mean	27.7		**7.9**	
Weaners on straw				
			7.2	Based on ratio slurry/straw for finishers

Table A3. Studies delivering poultry housing EF

Study	Emission g N $lu^{-1}d^{-1}$	No. studies	Emission Factor % TAN	Notes
Layers – deep-pit (cages, perchery, free-range)				
Peirson, 1995	79.0	3	22.1	Assume N excretion 0.82 kg (1995 value)
G Koerkamp, 1998	184.1	1	49.2	Assume N excretion 0.82 kg (1995 value)
G Koerkamp, 1998	146.1	1	39.0	Assume N excretion 0.82 kg (1995 value)
WA0368	139.2	1	36.8	Assume N excretion 0.79 kg (1998 value)
WA0651	196.8	1	57.9	Assume N excretion 0.78 kg (2000 value)
Mean	149.0		41.0	
Weighted mean	107.0		**35.6**	
Layers – deep litter: assume same EF as for perchery				
Layers – belt-cleaned (cages)				
Peirson, 1995	36.0	3	10.1	Assume N excretion 0.82 kg (1995 value)
WA0651 Gleadthorpe	79.2	1	23.3	Assume N excretion 0.78 kg (2000 value)
WA0651 comm. farm	64.8	1	19.1	Assume N excretion 0.78 kg (2000 value)
Mean	60.0		17.5	
Weighted mean	50.4		**14.5**	
Broilers				
Demmers et al. 1999	42.0	1	7.0	Assume N excretion 0.56 kg (1995 value)
Robertson et al 2002	44.0	4	8.3	Assume N excretion 0.55 kg (2000 value)
Frost et al 2002	54.0	4	9.2	Assume N excretion 0.55 kg (2000 value)
WA0651 winter	36.0	4	9.5	Derived N excretion from N balance
WA0651 summer	67.2	4	15.6	Derived N excretion from N balance
WA0651 drinkers	52.8	2	10.9	Derived N excretion from N balance
Mean	49.3	19	10.1	
Weighted mean	50.1		**10.5**	
Turkeys				
Peirson et al, 1995	93.0	3	**36.6**	

A measurement from Groot Koerkamp *et al*. (1998) for broiler housing (164 g N lu^{-1} d^{-1}) has been excluded from the inventory. This measurement was from a very old housing system, not representative of broiler housing, and was also based on a single measurement in time rather than an integrated measurement over the duration of the crop.

Table A4. Studies delivering cattle manure storage EF

Mean EF g N $m^{-2}d^{-1}$	Values g N $m^{-2}d^{-1}$	n	Emission as % TAN	Source
Slurry stores and lagoons without crusts				
3.42				Assumed to be double that for crusted stores (WA0641, WA0714)
Slurry stores and lagoons with crusts, weeping wall stores				
1.71	0.6		**2.3	(Phillips *et al*., in press)
	1.27, 3.65, 5.7		NA	WA0625
	0.44	2	*6.0	WA0632*
	1.8		NA	WA0641
	1.7		NA	Hill (2000)
	0.48	2	NA	WA0714
	0.5,0.72,0.42,0.7		51.5 (lagoons)	WA0717
	3		5.3 (w.wall)	AM0102
	4.2		NA	
Below ground slurry tanks				Assume same as for crusted above-ground tank
FYM heaps	**g N t^{-1} initial heap mass**			
265	421, 101, 106		NA	WA0618
		2	49	WA0519
		2	29	WA0632
		3	11	Chadwick, 2005
		2	31	WA0716
		1	11	Moral et al., 2012

** Emissions expressed per day. This value assumes 90 d storage.

Slurry stores are assumed to develop a crust unless they are stirred frequently.

Values derived from measurements made using Ferm tubes have been corrected to account for incomplete recovery of ammonia by Ferm tubes (Phillips *et al*., 1998). (*IGER values have been corrected using a factor of **0.7**).

Table A5. Studies delivering pig manure storage EF

Mean EF g N m^{-2} d^{-1}	Values g N m^{-2} d^{-1}	n	Emission as %TAN	Source
Slurry stores and lagoons				
3.16	1.34	4	13.0	WA0632
	2.47, 6.2		NA	WA0625
	2.4		NA	Phillips *et al*. (1997)
	1.56		NA	WA0708
	5.0		NA	Phillips *et al*. (1997)
Below ground slurry tanks				Assume 50% of EF for above-ground tank
FYM heaps	**g N t^{-1} initial heap mass**			
1224	539	4	20	WA0632
	1015	2	54	WA0716

Values derived from measurements made using Ferm tubes have been corrected to account for incomplete recovery of ammonia by Ferm tubes (Phillips *et al*., 1998).

Table A6. Studies delivering poultry manure storage EF

Mean EF	Values	n	Emission as %TAN	Source
g N t^{-1} initial heap mass				
Layer manure				
1956	318	2	3.5	WA0712
	3172	4	14.3	WA0651 (belt scraped)
	3141	1	29.5	WA0651 (deep pit)
	1193	1	20.0	WA0651 (belt scraped)
Litter				
1435	478	1	2.2	WA0712
	1949	4	19.9	WA0651 (winter)
	158	4	1.8	WA0651 (summer)
	639	2	8.4	WA0651 (drinkers)
	3949		NA	WA0716

Table A7: Studies delivering cattle and sheep grazing EF

	N input	Urine N	NH_3 emission	Due to fertiliser	Due to urine	Emission Factor
			Kg N ha^{-1}			%TAN
CATTLE						
Bussink	*Fert Res 33 257-265*					
1987	550	425	42.2	7.7	34.5	8
1988	550	428	39.2	7.7	31.5	7
1988	250	203	8.1	3.5	4.6	2
Bussink	*Fert Res 38 111-121*					
1989	250	64.2	3.8	3.5	0.3	0
1989	400	76.2	12.0	5.6	6.4	8
1989	550	94.3	14.7	7.7	7	7
1990	250	217.4	9.1	3.5	5.6	3
1990	400	339	27.0	5.6	21.4	6
1990	550	407.1	32.8	7.7	25.1	6
Lockyer	*J Sci Food Agric 35, 837-848*					
1	26	0.6455				2
2	26	0.7025				3
Jarvis et al	*J Ag Sci 112, 205-216*					
1986/87	0	69	6.7	0	6.7	10
1986/87	210	81	9.6	2.94	6.66	8
1986/87	420	207	25.1	5.88	19.22	9
AC0102						
Beef, North Wyke	0			0		10
Beef, Cambridge	0			0		7
SHEEP						
Jarvis et al	*J Ag Sci 117, 101-109*					
GC	0	169	1.1	0	1.1	1
HN	420	321	8.0	5.88	2.08	1
AC0102						
Boxworth	0					4
North Wyke	0					10

Table A8. Studies delivering EF for outdoor pigs

	Emission g N $lu^{-1}d^{-1}$	EF %TAN	Source
Outdoor sows/piglets	25	26.1	Williams et al. (2000)
	66*	NA	Welch (2003)

*This value is probably an overestimate as emission rates were below the detection limit on a number of occasions (and those data were not included).

The EF was derived from the Williams et al (2000) study, assuming the standard N excretion value for sows and a body weight of 200kg, giving a mean EF of 25 %TAN (assumed to be the same across all animal sub-categories).

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DEFRA Projects

Final reports from the following projects are available from Defra:

AC0114	GHG Platform – data management
AM0101	National ammonia reduction strategy evaluation system (NARSES)
AM0102	Modelling and measurement of ammonia emissions from ammonia mitigation pilot farms
AM0103	Evaluation of targeted or additional straw use as a means of reducing ammonia emissions from buildings for housing pigs and cattle
AM0110	Additional housing measurements for solid vs. liquid manure management systems
AM0111	Measurement and abatement of ammonia emissions from hard standings used by livestock
AM0115	Investigation of how ammonia emissions from buildings housing cattle vary with the time cattle spend inside them
DO108	Food and Agriculture Policy Research Institute – UK Project
ES0116	Field work to validate the manure incorporation volatilization system (MAVIS)
KT0105	Manure Nutrient Evaluation Routine (MANNER-NPK)
LK0643	UK Poultry Industry IPPC Compliance (UPIC)
NT2001	Integration of animal manures in crop and livestock farming systems: nutrient demonstration farms
NT2402	Impact of nutrition and management on N and P excretions by dairy cows
NT2605	The behaviour of some different fertiliser-N materials - Main experiments
OC9117	Ammonia emission and deposition from livestock production systems

WA0519	Enhancing the effective utilisation of animal manures on-farm through effective compost technology
WA0618	Emissions from farm yard manure based systems for cattle
WA0625	The effects of covering slurry stores on emissions of ammonia, methane and nitrous oxide
WA0632	Ammonia fluxes within solid and liquid manure management systems
WA0633	Predicting ammonia loss following the application of organic manures to land
WA0638	Low cost, aerobic stabilisation of poultry layer manure
WA0641	Low-cost covers to abate gaseous emissions from slurry stores
WA0651	Ammonia fluxes within broiler litter and layer manure management systems
WA0652	Field ammonia losses in sustainable livestock LINK Project LK0613
WA0653	Quantifying the contribution of ammonia loss from housed dairy cows to total N losses from dairy systems (MIDaS2)
WA0707	Effect of storage conditions on FYM composition, gaseous emissions and nutrient leaching during storage
WA0708	Covering a farm scale lagoon of pig slurry
WA0712	Management techniques to minimise ammonia emissions during storage and land spreading of poultry manures
WA0714	Natural crusting of slurry storage as an abatement measure for ammonia emission on dairy farms
WA0716	Management techniques to reduce ammonia emissions from solid manures
WA0717	Ammonia emissions and nutrient balance in weeping-wall stores and earth banked lagoons for cattle slurry storage
WA0720	Demonstrating opportunities of reducing ammonia emissions from pig housing
WA0722	Ammonia emission from housed dairy cows in relation to housing system and level of production
WT0715NVZ	Nitrogen and phosphorus output standards for farm livestock

Appendix 2
Reduction efficiencies for ammonia mitigation methods applicable to the UK ammonia emission inventory

Introduction

Agriculture is the major source of ammonia (NH_3) emissions to the atmosphere in the UK, accounting for >80% of anthropogenic emissions. Most of these emissions derive from urea excreted by farmed livestock (or uric acid in the case of poultry) and emissions will therefore arise wherever livestock excreta are deposited or managed i.e. at grazing, in livestock housing and during manure storage and application to land. Emissions also arise from inorganic nitrogen fertilisers applied to land. The emission factors used to quantify these emissions in the national inventory are reported separately. A growing number of potential mitigation methods applicable to one or more of the emission sources have been described in the literature. This report lists those that are currently included in the inventory of NH_3 emissions from UK agriculture together with the mean NH_3 emission reduction efficiency associated with each method. In addition, the current state of knowledge regarding the impact of the implementation of each method on emissions of nitrous oxide and methane is given so that these mitigation methods can be fully included in the revised combined agricultural GHG and NH_3 emission inventory.

Emission reduction methods

Only explicit mitigation methods are included here – i.e. those that are associated with a reduction in the emission factor for a particular source. Implicit mitigation methods, generally associated with efficiency improvements (e.g. a reduction in fertiliser use through better accounting for manure nitrogen use; a reduction in livestock numbers associated with productivity improvements), will be reflected in the inventory through changes in the activity data and are not described here. One exception in the current NH_3 emission inventory is the inclusion of a dietary measure, namely low crude protein diets for dairy cows, which is associated with a 20% reduction in the ammoniacal nitrogen content of dairy cow excreta over the housed winter period. In the revised emission inventories, N excretion will be derived using a balance approach according to diet and production characteristics and will therefore reflect any changes in the crude protein content of the diet.

Mitigation methods are categorised according to the emission source i.e. livestock housing, hard standings, manure storage, manure spreading and fertiliser application. Data sources are given, but the reported emission reduction efficiencies are not necessarily the arithmetic mean of reported studies but are more aligned with the expert judgement approaches used in the Defra 'Mitigation Methods - User Guide' (Newell Price et al., 2011) and the UNECE Task Force for Reactive Nitrogen 'Options for Ammonia Mitigation Guidance Document' (Bittman et al., 2014). These documents and other cited literature should be consulted for more detailed information on the mitigation methods included in Table 1.

Uncertainties are not well defined for these emission reduction estimates, so following 2006 IPCC Guidelines for Tier 2 approach to estimating emissions from manure management, uncertainty bound of ±20% of the reported value are applied with constraining limits of 0 and 100% also implemented.

Table 1. Reduction efficiencies for ammonia emission mitigation methods and an indication of their impact on nitrous oxide and methane emissions

Emission source	Mitigation method	Ammonia emission reduction efficiency (%)	Nitrous oxide	Methane	Data source
Cattle housing	Increased scraping frequency in cubicle house (from 2 to 4x per day)	15	-	-	Webb et al. (2006); Braam et al. (1997)
	Grooved flooring system for rapid urine draining	35	-	-	Swiestra et al. (2001); Bittman et al. (2014)
Pig housing	Partly slatted floor with reduced pit area	30	-	-	Bittman et al. (2014)
	Acid air scrubbing techniques	80	-	-	Bittman et al. (2014)
	Frequent slurry removal with vacuum system	25	-	-	Bittman et al. (2014)
	Floating balls on below-slat slurry surface	25	-	-	Bittman et al. (2014)
Poultry housing	Air drying of manure on laying hen manure belt systems	30	?	?	Bittman et al. (2014)
	Acid air scrubbing techniques	80	-	-	Bittman et al. (2014)
	Poultry litter drying (e.g. heat exchangers)	30	?	?	Defra WA0638
Dairy cow collecting yards	Wash down with water twice per day	70	-	-	Misselbrook et al. (2006)
Slurry storage	Crusting of cattle slurry	50	↑ EF from 0 to 0.005 (IPCC 2006)	↓ Methane Conversion Factor from 17 to 10% (IPCC 2006)	Misselbrook et al. (2005)
	Floating cover (e.g. expanded clay granules)	60	-	-	Bittman et al. (2014); Defra AC0115

	Tight lid, roof or tent structure	80	-	-	Bittman et al. (2014)
FYM/poultry manure storage	Sheeting cover	60	↓ by 30%	-	Chadwick (2005)
Slurry application	Trailing hose	30	-	-	Smith et al. (2000); Misselbrook et al. (2002); Bittman et al. (2014)
	Trailing shoe	60	-	-	Smith et al. (2000); Misselbrook et al. (2002); Bittman et al. (2014)
	Shallow injection	70	-	-	Smith et al. (2000); Misselbrook et al. (2002); Bittman et al. (2014)
Cattle slurry to arable	Incorporation within 4h by plough	59	-	-	Defra ES0116
	Incorporation within 4h by disc	52	-	-	Defra ES0116
	Incorporation within 4h by tine	46	-	-	Defra ES0116
	Incorporation within 24h by plough	21	-	-	Defra ES0116
	Incorporation within 24h by disc	19	-	-	Defra ES0116
	Incorporation within 24h by tine	17	-	-	Defra ES0116
Pig slurry to arable	Incorporation within 4h by plough	67	-	-	Defra ES0116
	Incorporation within 4h by disc	59	-	-	Defra ES0116
	Incorporation within 4h by tine	52	-	-	Defra ES0116
	Incorporation within 24h by plough	29	-	-	Defra ES0116
	Incorporation within 24h by disc	26	-	-	Defra ES0116
	Incorporation within 24h by tine	23	-	-	Defra ES0116
Cattle, pig and duck FYM	Incorporation within 4h by plough	71	-	-	Defra ES0116
	Incorporation within 4h by disc	47	-	-	Defra ES0116
	Incorporation within 4h by tine	39	-	-	Defra ES0116
	Incorporation within 24h by plough	34	-	-	Defra ES0116
	Incorporation within 24h by disc	23	-	-	Defra ES0116
	Incorporation within 24h by tine	19	-	-	Defra ES0116

Poultry manure	Incorporation within 4h by plough	82	-	-	Defra ES0116
	Incorporation within 4h by disc	64	-	-	Defra ES0116
	Incorporation within 4h by tine	45	-	-	Defra ES0116
	Incorporation within 24h by plough	56	-	-	Defra ES0116
	Incorporation within 24h by disc	44	-	-	Defra ES0116
	Incorporation within 24h by tine	31	-	-	Defra ES0116
Urea fertiliser	Urease inhibitor	70	?↓ (Smith et al. 2012)	-	Defra NT26
UAN fertiliser	Urease inhibitor	40	?	-	Defra NT26

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